

1 **Raising the bar for systematic conservation planning**

2

3 William T. Langford<sup>a\*</sup>, Ascelin Gordon<sup>a</sup>, Lucy Bastin<sup>b</sup>, Sarah A. Bekessy<sup>a</sup>,

4 Matt D. White<sup>c</sup>, Graeme Newell<sup>c</sup>

5

6 <sup>a</sup>School of Global Studies, Social Science & Planning, RMIT University, GPO Box 2476,

7 Melbourne, 3001 Australia.

8

9 <sup>b</sup>School of Engineering and Applied Science, Aston University, Birmingham, B4 7ET, UK

10 (currently at the Joint Research Centre of the European Commission, Ispra, Italy)

11

12 <sup>c</sup>The Arthur Rylah Institute for Environmental Research, Department of Sustainability and

13 Environment, PO Box 137, Heidelberg 3084, Australia.

14

15 *Corresponding author:* Langford, W.T. (bill.langford@rmit.edu.au)

16

17

18 **Abstract**

19

20 **Systematic conservation planning (SCP) represents a significant step toward cost-**

21 **effective, transparent allocation of resources for biodiversity conservation. However,**

22 **research demonstrates important consequences of uncertainties in SCP and of basing**

23 **methods on simplified circumstances involving few real-world complexities. Current**

24 **research often relies on single case studies with unknown forms and amounts of**

25 **uncertainty as well as low statistical power for generalizing results. Consequently,**  
26 **conservation managers have little evidence for the true performance of conservation**  
27 **planning methods in their own complex, uncertain applications. To build effective and**  
28 **reliable methods in SCP, we need more challenging and integrated testing of their**  
29 **robustness to uncertainty and complexity and much greater emphasis on generalization**  
30 **to real-world situations.**

31

## 32 **Systematic Conservation Planning: background and definition**

33 Widespread loss of biodiversity is commonly addressed by attempts to reserve, protect, and  
34 manage habitat for species at risk. Making good choices for these actions in a dynamic,  
35 spatial and temporal context is a difficult problem when there are many species, candidate  
36 actions, and locations involved. This leads to problems that are often not amenable to  
37 solution by rules of thumb or exactly solvable by analytic means [1].

38

39 Over the past 25 years, a family of mathematical approaches has evolved to explicitly define  
40 criteria and computationally solve for near-optimal prioritizations of conservation actions. A  
41 primary focus has been on mathematical methods for spatial allocation of conservation  
42 reserves [2-4] under the umbrella of Systematic Conservation Planning (SCP) [5-7]. In recent  
43 years, research in this area has also evolved toward a broader emphasis on prioritizing  
44 conservation actions in general through the lens of decision theory [8]. In this paper, we  
45 examine the collision between these high-precision methods and complex, highly uncertain  
46 data and discuss strategies for improving the utility and reliability of these methods.

47

48 The ideas behind SCP have much to offer conservation managers in moving beyond ad-hoc  
49 conservation planning, including the promise of quantitative, repeatable, and transparent  
50 decision-making. This is a significant advance given the inscrutable and idiosyncratic nature  
51 of conservation planning and investment across the globe. The resulting methods are  
52 powerful tools that have been used in numerous real-world conservation efforts with  
53 significant biodiversity implications, for example, selecting reserves for Madagascar [9-13],  
54 the Great Barrier Reef [14], and South Africa [15].

55

56 In spite of its successes and broad application, various authors have noted that conservation  
57 planning still encounters significant obstacles in bridging the gap from academic research to  
58 application. For example, a number of authors describe problems related to undertaking and  
59 implementing conservation planning within a complex web of social, economic, and political  
60 constraints[16-18]. Although these issues are at least as important, our focus here is on  
61 problems associated with mathematical aspects of SCP. Specifically, we discuss issues in  
62 addressing the effects of complexities and large uncertainties that are associated with most  
63 data and models used in SCP.

64

### 65 **Making conservation decisions under uncertainty**

66 While current SCP methods are mathematically sophisticated and highlight many important  
67 factors such as complementarity and risk, there are important mathematical difficulties in  
68 applying these results to real problems. In particular, much SCP research has focused on case  
69 studies conducted in simplified circumstances where most real-world uncertainties and  
70 complications are either poorly understood or abstracted away to make the problem  
71 mathematically tractable [19]. Unfortunately, uncertainties and approximations in data and  
72 models are ubiquitous in SCP, significantly affecting reliability of information about factors  
73 such as: costs and budgets; land availability; species vulnerabilities, presence, abundance,  
74 and interactions; as well as large-scale effects of climate, economics, land use change, and  
75 politics. These and numerous other complexities violate the assumptions of methods and  
76 effectively eliminate any theoretical guarantees of finding the best solution, or possibly even  
77 a good solution [20].

78

79 In recent years, numerous papers have explored this issue and demonstrated consequences of  
80 multiple types of uncertainties and complications as well as proposing methods for dealing

81 with some of them [20-36]. However, the conclusions (including our own work [20]) are only  
82 evaluated on case studies or problem sets with limited scope that provide little evidence for  
83 how they will carry over to different and more complex circumstances. More importantly,  
84 they do not provide explicit predictive models for the robustness of their conclusions to  
85 previously unseen situations and uncertainties. In fact, many studies carefully state that their  
86 results do *not* generalize beyond the current case.

87

88 While we raise questions here about the accuracy of SCP methods under uncertainty and  
89 complexity, not all errors are equally important. For a user making a decision, what matters is  
90 not just the exact amount of error in the output. Instead, the question is whether the decision  
91 and outcome would change if we could reduce the error, for example, by gathering more  
92 information [30]. Current SCP research would have much greater utility to practitioners if it  
93 could provide predictive models and evidence for estimating the amount, structure, and  
94 probability of error necessary for knowing whether decisions should change. In Box 1, we  
95 give an example of the consequences of this lack of predictive estimation.

96

97 In the rest of this paper, we suggest immediate solutions to parts of the problem and outline a  
98 research agenda to fill the gaps in building a better, more predictive understanding of the  
99 robustness of SCP methods that goes well beyond the scope of current studies. This  
100 understanding would in turn, lead to improving SCP methods and outcomes.

101

## 102 **Underlying problems**

103 Deciding whether the application of a given SCP method is acceptable in a particular  
104 situation requires knowledge of a) the errors in the inputs (i.e., the range and probability  
105 distribution of possible input errors), and b) how and when those errors are likely to affect the

106 outputs of the given SCP method. We believe that three specific issues significantly undercut  
107 a decision maker's ability to reliably estimate the distribution of these errors and their effects  
108 on predicted outcomes. These issues are: *Lack of predictive modelling and generalization,*  
109 *Unknown amounts of error, and Lack of good input error models.*

110

### 111 ***Lack of predictive modelling and generalization for SCP outputs***

112 The first fundamental problem is that the SCP literature relies almost exclusively on case  
113 studies or problem sets with very limited scope as opposed to studies across many types of  
114 problems leading to a testable prediction of the generality of the result. This means that most  
115 SCP research lacks statistical power to control for problem characteristics that drive the  
116 performance of SCP methods (Box 2). Varying these characteristics across studies is  
117 important because even applying SCP methods to the same landscapes, the same number of  
118 species, the same costs and the same input uncertainties can yield very different performance  
119 as a function solely of a single characteristic such as the structure of species distributions  
120 (Figure 1).

121

122 To our knowledge there are currently no studies that explicitly state and test a mathematical  
123 or verbal prediction of the size, probability, and spatial distribution of errors in SCP outputs  
124 on previously unseen problems. Predictions are not even made and tested over the typical  
125 range of the simplest and most ubiquitous variables such as the number of species and  
126 number of planning units. While many authors honestly state that their studies do not  
127 generalize, this means that there is little evidence to show that conclusions in specific case  
128 studies will hold in a user's own situation with its associated uncertainties and complications.

129

### 130 ***Unknown amounts of error: Apparent versus True values***

131 A second fundamental problem is that **it** is impossible to measure the true amount of error at  
132 any specific pixel, polygon, or map in both inputs and outputs of SCP studies that use real  
133 data. This is a result of the fact that data inputs to SCP are usually the result of a chain of  
134 sampling and model-fitting transformations. While we can improve the quality of SCP inputs  
135 through better field sampling, cost and time constraints mean that intensive sampling is often  
136 not an option. More importantly, for many inputs to SCP, it is impossible for sampling to  
137 identify the true value of that input. For example, even if we know exactly where every true  
138 location of a species is, an infinite number of distributions could have generated that sample  
139 (e.g., the uniform distribution). Consequently, when evaluating a conclusion, users must  
140 assume that input data are correct, meaning that they only see what we will refer to as the  
141 *apparent* performance of the techniques of interest (Figure 2).

142

143 This masking of the *true* performance means that studies comparing methods or rules of  
144 thumb on *apparent* data may provide meaningless and/or misleading results. For example,  
145 many papers emphasize the importance of including costs in conservation planning [10, 37-  
146 40], but those conclusions are based on *true* costs. What do they tell us about using *apparent*  
147 costs in real cases where the values used to compute these measures are all uncertain? Adams  
148 et al. [41] give an example that raises the question of whether “using partial estimates of costs  
149 might produce less efficient outcomes than assuming homogeneous costs”. While there is no  
150 question about the usefulness of *correct* costs in choosing actions, how do users like those in  
151 the Madagascar example (Box 1) make sure their decisions are robust to errors in costs? The  
152 problem with ignoring cost is that it implicitly assumes homogeneous costs, and this usually  
153 represents a large amount of error with respect to the true costs. This suggests that questions  
154 about ignoring costs in SCP are better seen as questions about the consequences of *error* in  
155 cost. Therefore, just as with species data, we need to pay attention to determining when the

156 use of apparent values rather than true values will lead to a change of decision.

157

158 ***Lack of good error models for SCP inputs***

159 The third problem is that even when we can know true error values, we have little evidence  
160 for the correlation structure and distribution of the set of errors (Figure 2). Both sensitivity  
161 analysis and efforts at generalization under uncertainty rely on the correctness of the *error*  
162 *models* assumed to generate or describe the amount, structure, and probability of errors in  
163 inputs to the SCP process. However, there is currently so little testing of inputs to SCP  
164 methods with known amounts and structures of error that there is very little evidence for  
165 these generative error models. Instead, errors or perturbations are generated based on opinion  
166 rather than evidence, in spite of the significant literature reflecting the common  
167 overconfidence of experts in their own opinions [42-44]. Given the spatial, sequential and  
168 combinatorial nature of SCP, an important missing aspect of error models is the spatial and  
169 temporal specificity and interactions of different types of errors.

170

171 **Recommendations**

172 We believe that several steps can be taken to make it easier for SCP users to know the risks  
173 and rewards associated with a particular decision. These steps do not require new  
174 mathematical techniques; rather, they require a shift in the culture and orientation of SCP  
175 research. In short, we believe that large improvements can be made simply by moving away  
176 from the descriptive, anecdotal, case study culture to one that is predictive and evidence-  
177 based, striving to support generalizable results. The essence of the strategy we support is  
178 simply to augment the many small-scale studies that currently exist by testing explicit,  
179 predictive claims about the performance and error in SCP methods on a statistically powerful  
180 variety of problems, using data with known amounts of error and controlling for problem

181 complexity and uncertainty. We give more detail below and provide a summary in Figure 2.

182

183 *Explicitly predict probability, structure, and amount of SCP output error and test*

184 *generalization of predictions*

185 We need to learn predictive models that map problem characteristics to probability

186 distributions of errors in SCP output. We also need to make their predictions explicit and test

187 them on problems that cover the space of SCP problems and input errors and where we can

188 know the true answer. The space that we refer to here includes factors such as number of

189 species, number of planning units, correlations of species co-occurrence, species rarity, cost,

190 etc.

191

192 As a simple example, we would like to be able to predict the following: given the

193 environmental, economic, and ecological samples used in solving a specific reserve selection

194 problem using apparent data, what are the confidence intervals on the estimated i) true

195 representation achieved, ii) true cost of achieving representation goals, and iii) the decrease in

196 outcome quality resulting from using apparent rather than true data (also known as the

197 *regret*). We would actually like to know the distributions of these errors as well as the

198 downstream effects that we really care about such as persistence or vegetation condition. For

199 now though, even simple aggregate measures like variance or a confidence interval would be

200 a useful first step.

201

202 *Challenge problems*

203 Faced with problems of similarly daunting complexity, ecologists often conduct controlled

204 experiments where they create their own communities and manipulate them to gain

205 information about real systems. We suggest that in a similar way, we can build families of

206 challenge problems that act as surrogate systems to understand and learn to predict the  
207 reliability of SCP methods. While we do not understand the world well enough to build a  
208 model of it that mirrors its complex and stochastic behavior, we do understand many drivers  
209 that lead to complex behavior in environmental and economic systems. With existing  
210 knowledge and techniques, we can generate families of similarly dynamic, complex, and  
211 plausible situations spanning the space of SCP problems.

212

213 Linking multiple generative models to build families of SCP problems of varying difficulty  
214 and complexity would give us both true and apparent data as well as sufficient statistical  
215 power and control for training and for independent test examples to estimate generalization.  
216 Having control over problem structure would also make it possible to build benchmark  
217 problems of varying difficulty, which would allow tests to screen for the relative power of  
218 different methods. While this is not sufficient to predict exactly how conclusions will hold in  
219 new situations, one necessary condition for a useful conclusion is that it hold on simpler  
220 problems than the ones where we want to apply it.

221

222 *Active learning*

223 One problem with learning to predict is that there is a huge range of problem structures and  
224 uncertainties to consider. One way to make this more tractable is to use techniques such as  
225 active learning to reduce the total number of experiments required. The intuition here is that  
226 we need not explore every region of the problem space with equal effort if we are doing well  
227 in certain parts of the space with little sampling. In active learning, an algorithm asks for new  
228 samples as it learns rather than specifying the complete sample set beforehand as might be  
229 done using optimal experimental design. The choice of the next sample can then be based on  
230 where the algorithm most needs information to improve its predictions. This may save

231 considerable resources compared to preselecting all sample points before simulation.

232

233 *Explicitly seek and test for robustness to unknown uncertainties*

234 Since we can never know all of the factors and uncertainties to include in our models, our

235 goal should be to mitigate this kind of model error by designing and demonstrating

236 robustness to these unknown factors. By definition, we cannot know what these unknown

237 unknowns are. However, the current literature has already given us a large catalog of

238 uncertainties that we can use to explore and expand the robustness of methods and beliefs.

239 We can do this in much the same way that we use bootstrap sampling to estimate the variance

240 of samples where the true distribution is unknown. For example, if a method or proposition

241 is tested on a number of uncertainties that it has made no attempt to address, then that gives

242 us information on its robustness to unseen uncertainties.

243

244 *Measure results on data with known amounts of error*

245 The only way to know a method's true performance under uncertainty (as opposed to its

246 apparent performance), is to generate the errors ourselves and apply them to data that we

247 define or generate as correct (Figure 2). Clearly, the results of simulation are only as useful as

248 their match to reality, but it is the only way that we can accurately measure the consequences

249 of error given our limited ability to sample and experiment on the real world. It is important

250 to note that while "real" data may appear to be more valid than simulations, studies that use

251 real data still include unknown amounts of error and therefore, have the same problems with

252 being approximations to reality as modelled results. Moreover, any real data can be used in a

253 simulation as well, so simulations represent a superset of what can be done solely with real

254 data. Simulation is not the only answer to all problems in SCP, but it is a necessary adjunct

255 to field data by providing another experimental environment for collecting evidence about the

256 behavior of methods under uncertainty [45-48].

257

258 ***Learn error models predicting probability, structure, and amount of input errors for SCP***

259 To perform sensitivity analysis on SCP outputs we must learn models to generate the  
260 magnitude and distribution of errors that occur in inputs to SCP such as habitat suitability and  
261 cost models. For example, if errors in cost or species input maps are distributed uniformly  
262 across a study area, they will have different consequences for the results of a spatially explicit  
263 reserve selection algorithm than if the errors are spatially correlated with factors like soil type  
264 and patch boundaries (Figure 2).

265

266 We can develop much better error models than are currently available simply by applying  
267 known uncertainties and biases to input data and then propagating those effects through the  
268 SCP process [48], as described in Challenge Problems above. As a simple example, we  
269 know that there is often sampling bias in inputs to habitat models. This bias is easy to  
270 simulate and propagate through the habitat modelling. We can then test the robustness of SCP  
271 to this kind of error. If we know distributions, biases, and magnitudes of these errors, then we  
272 have evidence for choosing bounds or distributions of scenarios to test as well as the ability  
273 to investigate the value of gathering more information to reduce uncertainty [29-30].

274

275 ***Raising the bar***

276 Finally, editors, reviewers, and funding agencies would aid practitioners by insisting that  
277 research goes beyond case studies, mathematical novelty, and the impact of a single type of  
278 complexity or uncertainty in a single distribution in a single location. Ginzburg and Jensen  
279 have made a similar point in relation to theoretical ecology [49]:

280

281 “An engineering firm that builds a faulty bridge based on an overfitted model will be  
282 sued or fined out of existence; to date, we know of no ecological theorist whose  
283 similarly overfitted model has evoked comparable penalties. Because society  
284 demands little from theoretical ecology, one can have a successful lifetime career in  
285 the field without any of one’s theories being put to the practical test of actual  
286 prediction.”

287

288 True progress in SCP methods and outcomes requires a culture that expects new studies to  
289 make and test predictions and to control for the structure of problems and methods known to  
290 affect performance.

291

## 292 **Conclusion**

293 SCP is undoubtedly a useful and important advance in conservation decision-making.  
294 However, for SCP to be truly useful for conservation managers, the reliability of its outputs  
295 must be honestly characterized. No matter whether a conservation plan is the result of a  
296 negotiation or the output of an algorithm, we need to know: i) the risks and rewards of  
297 attempting to use the proposed method or strategy in a particular location and ii) the  
298 probability of it achieving its claimed outcomes.

299

300 SCP will benefit greatly if we raise the bar and undertake research that builds predictive  
301 models and evidence for method performance rather than over-simplified case studies on data  
302 containing unknown amounts of error. The more rigorous quantification of problem  
303 characteristics and error that we propose will also create potential for systematic meta-  
304 analysis and meaningful comparison of results.

305

306 Some may claim that performance prediction is in fact impossible because the real world is  
307 too complex, but our central concern is that it cannot go both ways. *If the problem is so*  
308 *complex that reliable bounds on a proposed SCP method's performance in a new, unseen*  
309 *situation are impossible, then it is also impossible to claim a priori that the proposed method*  
310 *will give reliable performance in new unseen situations.*

311

312 Given that SCP outcomes may determine the fate of some species, addressing this problem is  
313 not “just academic”.

314

### 315 **Acknowledgements**

316 We would like to thank Iris Bergmann, Ben Cooke, Alex Lechner, Katelyn Samson, and two  
317 anonymous reviewers for their helpful comments on the manuscript. This research was  
318 funded by the Australian Research Council through Linkage Project LP0882780, and the  
319 Applied Environmental Decision Analysis research hub (through the Australian  
320 Commonwealth Environment Research Facilities program).

321

322

323 **Box 1 – What is the evidence for a plan’s likely performance? A large-scale example**

324

325 An exchange in the journal *Science* in 2008 illustrates the dilemma faced by potential users of  
326 SCP methods. Kremen et al. [9] used Zonation reserve selection software [50] to determine  
327 the best 10% of Madagascar to reserve for biodiversity conservation, based on data from six  
328 taxonomic groups. Later, Coetzee [12] criticized the plan for not incorporating costs, well-  
329 known taxa (e.g., birds), and effects of climate change. Bode et al. [10] further criticized the  
330 exclusion of costs from the prioritization, claiming that the plan included expensive areas and  
331 suggesting that a better solution would favor low cost areas. In reply, Kremen et al. [11]  
332 pointed out that the costs in Bode et al. were based on global data that was inaccurate, at the  
333 wrong scale, and based solely on agriculture and livestock opportunity costs. Furthermore,  
334 they had addressed taxonomy and climate change through their choice of species and were  
335 unable to get appropriate bird data for their procedures.

336

337 Two years later, Fiorella et al. [13] applied Zonation to two different versions of a subset of  
338 the Madagascar data (52 lemur species). As in range studies with other species [51-54], they  
339 found large overestimation of habitat in the original maps. After using species knowledge to  
340 restrict the maps and re-running Zonation they found that 65% to 68% of locations that were  
341 in the original top 10% (i.e., areas to be reserved) were no longer in the top 10%. This  
342 suggests that the original study could contain large errors, but SCP users still have no  
343 evidence for how likely that is, since the input maps for Zonation in the original study were  
344 generated using a different method.

345

346 Unfortunately for conservation managers needing to know the consequences of following the

347 plan, both the discussion and the research conclude with an unresolved exchange of opinions.  
348 Even given the entire SCP literature, there is no evidence or predictive model that would  
349 allow either side to put mathematically supported arguments for whether the decisions should  
350 be changed.

351

352 Our intention here is not to single out this work as problematic. Rather, we chose it because  
353 of its importance and the unique exchange of letters and follow-on study. Taken together,  
354 they encapsulate the state of SCP and point out that nearly all work in SCP, even high profile,  
355 expert work such as this, stops short of giving users the evidence they need to estimate the  
356 decision consequences of their uncertainty in the real world.

357

358

359

360 **Box 2 - Generalization and problem characterization**

361

362 When a practitioner with a conservation task looks to the SCP literature, several questions  
363 arise in determining whether the results apply to their situation. In particular, they need to  
364 know which studies:

- 365 - make explicit performance predictions and test them on previously unseen problems;
- 366 - use data where true outcomes can be known;
- 367 - use broad problem sets that give statistical power and are representative of real  
368 applications (multiple distributions, costs, threats, etc.);
- 369 - explore multiple uncertainties simultaneously;
- 370 - include correlation and spatial structure in the errors..

371

372 Numerous studies address single uncertainties but only use case studies lacking problem  
373 characterization to give users evidence for their situation. Below we describe four studies that  
374 take positive steps toward characterizing problem structure and performance bounds, though  
375 there are others that take positive steps as well (see [20,27-29] for other examples).

376

- 377 - A good example of putting rigorous bounds on expected performance under an uncertainty  
378 is given by Moilanen et al. [23]. Here, a worst-case analysis is used to guarantee a lower  
379 bound on species representation in outputs assuming a given worst-case input map error. This  
380 is a positive step but it is still conditioned on having a good input error model, for which we  
381 currently lack evidence. It also does not address *regret*, that is, the loss in method  
382 performance compared to what would have been achieved using the true data.

383

384 - Pressey et al. [24] address problem characterization by demonstrating variation in outcomes  
385 resulting from changes of data set size, site size, rarity of features, and nestedness of features  
386 in replicated synthetic data sets. All the data sets, however, were variations derived from a  
387 single original data set, consequently only exposing outcomes from one corner of the  
388 problem domain.

389

390 - Drechsler [25] addresses uncertain dynamics of land acquisition through simulations  
391 synthesizing different combinations of species counts, occupancy levels, and nestedness.  
392 While there is explicit uncertainty in the land acquisition, the method assumes that  
393 probabilities are known and correct, which is unlikely and known to bring other risks [55].

394

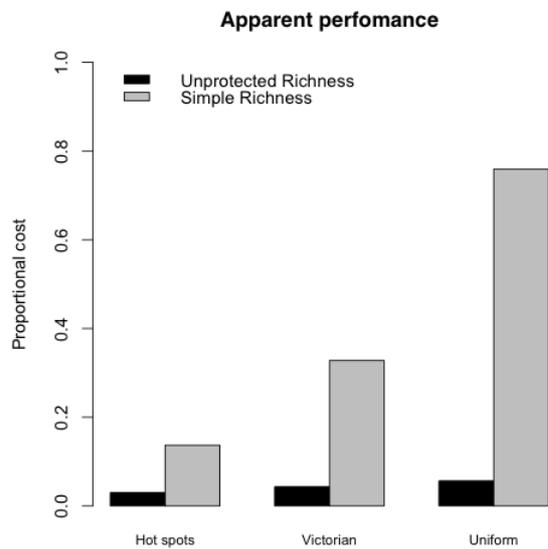
395 - Turner and Wilcove [26] also examine uncertainty in site availability, this time with a ten  
396 year time frame and three different real-world data sets, budget constraints, and loss of sites  
397 to development. However, they do not characterize the structure of the species sets to control  
398 for those effects and they ignore other complications, including uncertainty in the species  
399 data.

400

401

402 **Figure 1 - Problem difficulty and *Apparent vs. True* performance**

403



404

405 This figure presents results from [20] illustrating both the difficulties of relying on case  
406 studies and the utility of using simulated data to examine method performance under  
407 uncertainty.

408

409 The objective was to find the least-cost reserve network that contains at least one  
410 representation of each species using two common reserve selection heuristics based on  
411 species richness. The "simple richness" rule picks patches in order according to the number  
412 of species on each patch. "Unprotected richness" picks according to the number of species on  
413 each patch that have not already been reserved (complementarity). Cost is measured as a  
414 proportion of total landscape cost.

415

416 Error was introduced in the form of a 30% overestimate of suitable habitat since  
417 overestimation is common [51-54]. The results are shown for three different distributions of

418 species richness:

419 - “Hot spots”, where species tend to co-occur on the same patch;

420 - “Victorian”, where the distribution of co-occurrences matches a real distribution

421 in Victoria, Australia;

422 - “Uniform”, where species locations are uncorrelated.

423

424 The bar chart on the left shows the *apparent* costs of a reserve network that *appears* to

425 represent each species at least once based on the erroneous maps. The bar chart on the right

426 shows the *true* costs required. The dashed box around the *apparent* Victorian results

427 highlight what would have been found in a single case study using "real" data.

428

429 These results illustrate four important points discussed throughout this paper:

430

431 1. *Problem difficulty*: Even though the landscapes were identical and the number of patches

432 occupied by each species was identical, controlling for the spatial distribution of the species

433 revealed a wide range of costs from both methods.

434

435 2. *Misleading ranking of methods*: Based on *apparent* data, the Unprotected Richness

436 approach appears to be far more efficient than the Simple Richness approach, even though

437 the results were approximately equal on the *true* data.

438

439 3. *Misleading performance measurements*: Using the *apparent* maps alone, one would

440 grossly underestimate the cost required to achieve the conservation goal using the

441 Unprotected Richness approach. Even though Simple Richness never appeared to perform as

442 well as Unprotected Richness, its *apparent* performance was always similar to its *true*

443 performance.

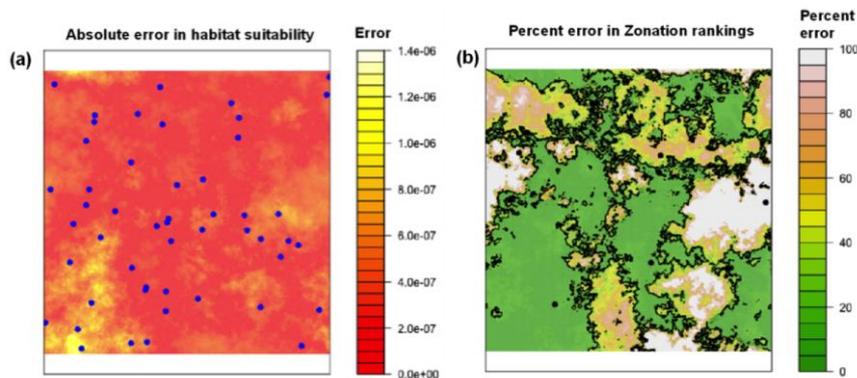
444

445 4. *Error amplification*: Results here and in [20] show how SCP methods can strongly  
446 amplify their input errors: *apparent* input habitat area is only 1.3 times *true* area, but *true*  
447 costs for Unprotected Richness were 4.6, 6.2, and 14.1 times their *apparent* costs.

448

449

450 **Figure 2 - Summary of specific suggestions for improving SCP studies under**  
451 **uncertainty**



452

453 Here we summarize specific recommendations for improving SCP studies in the presence of  
454 uncertainty and real-world complications.

455

456 1. **Building and testing predictive models of error** Explicitly state predictions for the  
457 probabilities and structure of error for any study's conclusions. Test predictions on  
458 multiple problems including interacting uncertainties on out-of-sample data. The two  
459 images above illustrate the kind of spatial correlation in error that must be accounted  
460 for. The left-hand image shows error in output of a habitat suitability model with  
461 simulated species presence locations. The right-hand image shows percent difference  
462 in priority rankings for a reserve selection method (Zonation) between using true  
463 suitability probabilities and corresponding erroneous probabilities from the habitat  
464 suitability model.

465

466 2. **Evaluation** When evaluating or comparing methods or studying uncertainties the  
467 following 3 steps should be included: (i) Measuring performance on data where true  
468 values are known, rather than on data containing unknown amounts of error; (ii)  
469 Measuring performance under many uncertainties simultaneously, rather than one at a

470 time; (iii) Characterization of problem structure (e.g. the number and spatial co-  
471 occurrence of species, landscape configuration, distribution of costs) and measuring  
472 performance on multiple problems with different structures.

473

474 3. **Sensitivity Analysis** Sensitivity analyses should be done by simultaneously varying  
475 multiple factors rather than studying each factor in isolation. Evidence should be  
476 given supporting the range and structure of errors used in the sensitivity analysis.

477

478 To make the method of measuring true versus apparent performance more concrete, we  
479 outline steps for this approach below (more details and examples are given in [20]):

480

481 • First, we define a dataset to be the true conditions in our modelled world. This data  
482 may be synthetic or real, and could represent any or all of the inputs to an SCP  
483 problem such as species habitat maps, cost maps, etc. With synthetic data, direct  
484 control over the input data characteristics is possible, or real-world data can be used  
485 from multiple locations where problem characteristics differ.

486

487 • Second, we can inject error into these true inputs to create apparent inputs by applying  
488 models representing our beliefs about real-world forms of error, for example, over or  
489 under-estimation of costs or spatial bias in error in predicting species habitat.

490

491 • Third, SCP analysis can then be carried out in parallel on both true and apparent data.  
492 The impact of errors being studied can then be determined by comparing differences  
493 between the SCP outcomes in the true and degraded data and predictive models can  
494 be built.

495 **References**

- 496 1 Church, R. (1996) Reserve selection as a maximal covering location problem. *Biol*  
497 *Conserv* 76, 105-112
- 498 2 Williams, J.C., *et al.* (2005) Spatial attributes and reserve design models: a review.  
499 *Environ Model Assess* 10, 163-181
- 500 3 Sarkar, S., *et al.* (2006) Biodiversity Conservation Planning Tools: Present Status and  
501 Challenges for the Future. *Annu Rev Env Resour* 31, 123-159
- 502 4 Vanderkam, R.P.D., *et al.* (2007) Heuristic algorithms vs. linear programs for designing  
503 efficient conservation reserve networks: Evaluation of solution optimality and processing  
504 time. *Biol Conserv*, 137, 349-358
- 505 5 Kirkpatrick, J. (1983) An iterative method for establishing priorities for the selection of  
506 nature reserves: An example from Tasmania. *Biol Conserv* 25, 127-134
- 507 6 Margules, C.R., and Pressey, R.L. (2000) Systematic conservation planning. *Nature* 405,  
508 243-254
- 509 7 Pressey, R.L., *et al.* (1993) Beyond opportunism: Key principles for systematic reserve  
510 selection. *Trends Ecol Evol* 8, 124-128
- 511 8 Wilson, K.a., *et al.* (2009) Setting conservation priorities. *Ann NY Acad Sci* 1162,  
512 237-264
- 513 9 Kremen, C., *et al.* (2008) Aligning conservation priorities across taxa in Madagascar with  
514 high-resolution planning tools. *Science* 320, 222-226
- 515 10 Bode, M., *et al.* (2008) The Cost of Conservation. *Science* 321, 340
- 516 11 Kremen, C., *et al.* (2008) Response to Conservation with Caveats. *Science* 321, 341-342
- 517 12 Coetzee, B.W.T. (2008) Conservation with Caveats. *Science* 321, 340-341
- 518 13 Fiorella, K., *et al.* (2010) Methodological considerations in reserve system selection: A  
519 case study of Malagasy lemurs. *Biol Conserv* 143, 963-973

- 520 14 Fernandes, L., *et al.* (2005) Establishing Representative No-Take Areas in the Great  
521 Barrier Reef: Large-Scale Implementation of Theory on Marine Protected Areas.  
522 *Conserv Biol* 19, 1733-1744
- 523 15 Cowling, R. (2003) A conservation plan for a global biodiversity hotspots in the Cape  
524 Floristic Region, South Africa. *Biol Conserv* 112, 191-216
- 525 16 Prendergast, J.R., *et al.* (1999) The Gaps between Theory and Practice in Selecting  
526 Nature Reserves. *Conserv Biol* 13, 484-492
- 527 17 Knight, A.T., *et al.* (2006) Designing Systematic Conservation Assessments that Promote  
528 Effective Implementation: Best Practice from South Africa. *Conserv Biol* 20,  
529 739-750
- 530 18 Milner-Gulland, E.J., *et al.* (2010) Do we need to develop a more relevant conservation  
531 literature? *Oryx* 44, 1-2
- 532 19 Moilanen, a. (2008) Two paths to a suboptimal solution – once more about optimality in  
533 reserve selection. *Biol Conserv* 141, 1919-1923
- 534 20 Langford, W.T., *et al.* (2009) When do conservation planning methods deliver ?  
535 Quantifying the consequences of uncertainty. *Ecol Inform* 4, 123-135
- 536 21 Regan, H.M., *et al.* (2002) A Taxonomy and Treatment of Uncertainty for Ecology and  
537 Conservation Biology. *Ecol Appl* 12, 618-628
- 538 22 Ascough II, J., *et al.* (2008) Future research challenges for incorporation of uncertainty in  
539 environmental and ecological decision-making. *Ecol Model* 219, 383-399
- 540 23 Moilanen, A., *et al.* (2006) Uncertainty analysis for regional-scale reserve selection.  
541 *Conserv Biol* 20, 1688-1697
- 542 24 Pressey, R.L., *et al.* (1999) Effects of data characteristics on the results of reserve  
543 selection algorithms. *J Biogeogr* 26, 179-191
- 544 25 Drechsler, M. (2005) Probabilistic approaches to scheduling reserve selection. *Biol*

- 545        *Conserv* 122, 253-262
- 546    26 Turner, W.R., and Wilcove, D.S. (2006) Adaptive decision rules for the acquisition of  
547        nature reserves. *Conserv Biol* 20, 527-537
- 548    27 Costello, C., and Polasky, S. (2004) Dynamic reserve site selection. *Resour Energy*  
549        *Econ* 26, 157-174
- 550    28 Moilanen, A., and Cabeza, M. (2007) Accounting for habitat loss rates in sequential  
551        reserve selection: Simple methods for large problems. *Biol Conserv* 136, 470-482
- 552    29 Visconti, P., *et al.* (2010) Habitat vulnerability in conservation planning—when it matters  
553        and how much. *Conserv Lett* 3, 404-414
- 554    30 Polasky, S., and Solow, A.R. (2001) The value of information in reserve site selection.  
555        *Biodivers Conserv* 10, 1051-1058
- 556    31 Polasky, S., *et al.* (2000) Choosing reserve networks with incomplete species  
557        information. *Biol Conserv* 94, 1-10
- 558    32 Polasky, S., *et al.* (2005) Conserving Species in a Working Landscape : Land Use with  
559        Biological and Economic Objectives. *Ecol Appl* 15, 1387-1401
- 560    33 Rae, C., *et al.* (2007) Implications of error and uncertainty for an environmental planning  
561        scenario: A sensitivity analysis of GIS-based variables in a reserve design exercise.  
562        *Landscape Urban Plan* 79, 210-217
- 563    34 Beech, T., *et al.* (2008) A stochastic approach to marine reserve design: Incorporating  
564        data uncertainty. *Ecol Inform* 3, 321-333
- 565    35 Nicholson, E., and Possingham, H.P. (2007) Making conservation decisions under  
566        uncertainty for the persistence of multiple species. *Ecol Appl* 17, 251-265
- 567    36 Leroux, S.J., *et al.* (2007) Accounting for System Dynamics in Reserve Design.  
568        *Ecol Appl* 17, 1954-1966
- 569    37 Ando, A., *et al.* (1998) Species Distributions, Land Values, and Efficient Conservation.

570 *Science* 279, 2126-2128

571 38 Shogren, J.F., *et al.* (1999) Why economics matters for endangered species protection.  
572 *Conserv Biol* 13, 1257-1261

573 39 Naidoo, R., *et al.* (2006) Integrating economic costs into conservation planning. *Trends*  
574 *Ecol Evol* 21, 681-687

575 40 Naidoo, R., and Iwamura, T. (2007) Global-scale mapping of economic benefits from  
576 agricultural lands: Implications for conservation priorities. *Biol Conserv* 140, 40-49

577 41 Adams, V.M., *et al.* (2010) Opportunity costs: Who really pays for conservation?  
578 *Biol Conserv* 143, 439-448

579 42 Henrion, M., and Fischhoff, B. (1986) Assessing uncertainty in physical constants.  
580 *Am J Phys* 54, 791-798

581 43 Kynn, M. (2008) The ‘heuristics and biases’ bias in expert elicitation. *J Roy Stat Soc A*  
582 *Sta* 171, 239-264

583 44 Speirs-Bridge, A., *et al.* (2010) Reducing overconfidence in the interval judgments of  
584 experts. *Risk Anal* 30, 512-523

585 45 Zurell, D. *et al.* (2010) The virtual ecologist approach: simulating data and observers  
586 *Oikos* 119, 622–635

587 46 Austin, M.P., *et al.* (2006) Evaluation of statistical models used for predicting plant  
588 species distributions: Role of artificial data and theory. *Ecol Model*, 199, 197-216

589 47 Meyer, K.M., *et al.* (2009) The power of simulating experiments. *Ecol Model*  
590 220, 2594-2597

591 48 Hoffman, J.D., *et al.* (2010) Use of simulated data from a process-based habitat model to  
592 evaluate methods for predicting species occurrence. *Ecography* 33, 656-666

593 49 Ginzburg, L.R., and Jensen, C.X.J. (2004) Rules of thumb for judging ecological theories.  
594 *Trends Ecol Evol* 19, 121-126

595 50 Moilanen, A., *et al.* (2005) Prioritizing multiple-use landscapes for conservation: methods  
596 for large multi-species planning problems. *P Roy Soc B-Biol Sci* 272, 1885-1891

597 51 Jetz, W., *et al.* (2008) Ecological Correlates and Conservation Implications of  
598 Overestimating Species Geographic Ranges. *Conserv Biol* 22, 110-119

599 52 Hurlbert, A.H., and Jetz, W. (2007) Species richness, hotspots, and the scale dependence  
600 of range maps in ecology and conservation. *P Natl Acad Sci USA* 104, 13384

601 53 Kelvey, K.S., *et al.* (2008) Using Anecdotal Occurrence Data for Rare or Elusive Species:  
602 The Illusion of Reality and a Call for Evidentiary Standards. *BioScience* 58, 549-555

603 54 Palminteri, S., *et al.* (2009) Usefulness of species range polygons for predicting local  
604 primate occurrences in southeastern Peru. *Am J Primatol* 9, 1-9

605 55 Regan, H.M., *et al.* (2005) Robust Decision-Making Under Severe Uncertainty for  
606 Conservation Management. *Ecol Appl* 15, 1471-1477